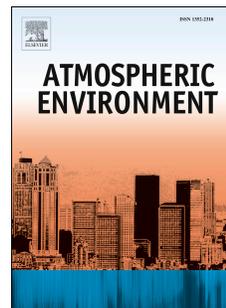


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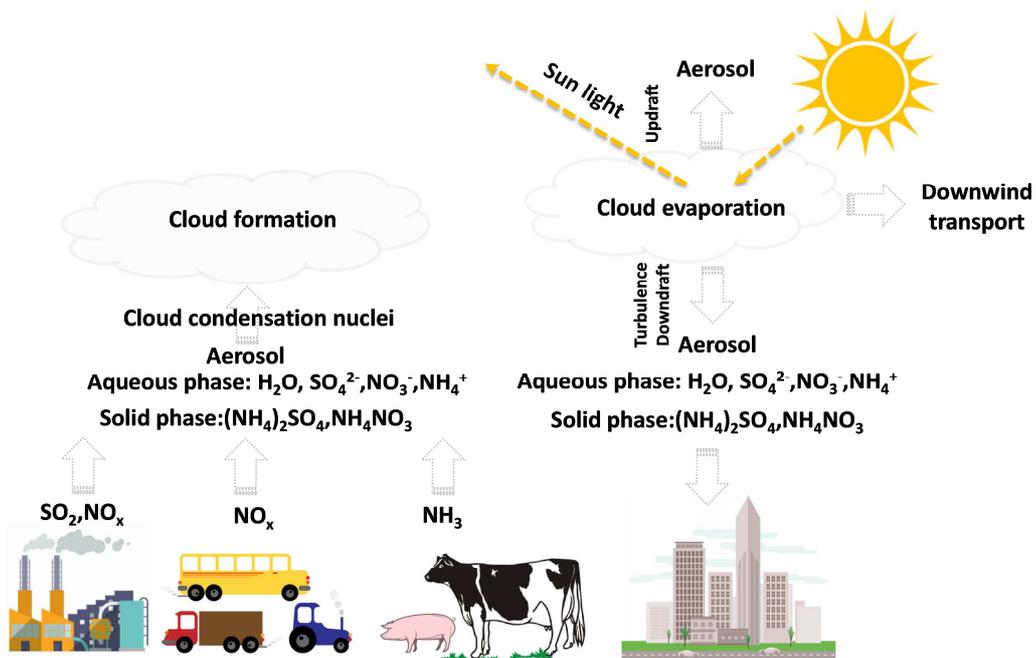
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## Observational evidence of cloud processes contributing to daytime elevated nitrate in an urban atmosphere

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1 **Abstract.** Nitrate has become the most important hydrophilic chemical component in  
2  $PM_{2.5}$  during serious air pollution periods in urban areas of the Pearl River Delta  
3 (PRD) region of south China, but there is a lack of fully understanding of its  
4 formation mechanisms and controlling factors, especially during daytime nitrate  
5 episodes. To fill the knowledge gap, water-soluble inorganic ions in  $PM_{2.5}$  and trace  
6 gases including  $SO_2$ ,  $HNO_3$ ,  $NH_3$ ,  $NO_x$  and  $O_3$ , particle size distribution, vertical  
7 profile of aerosol backscatter density, and ground-level and vertical profiles of key  
8 meteorological variables were simultaneously measured at high time resolution in  
9 urban Guangzhou of the PRD. The remarkably enhanced nitrate mass concentrations  
10 observed at the surface-level during daytime were identified to be caused by strong  
11 boundary-layer turbulent mixing of residual aerosols produced from evaporation of  
12 low-level shallow stratocumulus clouds. This finding may have important  
13 implications on the sources of secondary inorganic aerosols in this and other similar  
14 cloudy regions.

15

16 Key words: nitrate episode; aerosol size distribution; cloud processed aerosols;  
17 boundary-layer mixing

18

## 19 1. Introduction

20 Nitrate constitutes a substantial fraction of fine particles and thus has significant  
21 impacts on air quality and climate related issues such as acid deposition, reduced  
22 visibility, radiative forcing, and aerosol-cloud interaction (*Shindell et al., 2009; Tao et*  
23 *al., 2017; Watson, 2002; Zhang et al., 2007*). Nitrate is one of the most important  
24 pathways removing nitrogen oxides (NO<sub>x</sub>), which are the major primary emissions  
25 produced from fossil fuel combustion processes, soil, biomass burning and lightning,  
26 from the atmosphere (*Andreae & Merlet, 2001; Davidson & Kingerlee, 1997;*  
27 *Hudman et al., 2007; Richter et al., 2005*). In the atmosphere, NO<sub>x</sub> can be converted  
28 into nitrate in two chemical mechanisms (*Zhang et al., 2015*). The first one is through  
29 the oxidation of nitrogen dioxide (NO<sub>2</sub>) to HNO<sub>3(g)</sub> with the media of hydroxyl radical  
30 (OH·) during daytime. The second one is through the chemical reaction between NO<sub>2</sub>  
31 and O<sub>3</sub> to form nitrate radical (NO<sub>3</sub>), which further reacts with NO<sub>2</sub> to form dinitrogen  
32 pentoxide (N<sub>2</sub>O<sub>5</sub>) during nighttime. HNO<sub>3(g)</sub> and (N<sub>2</sub>O<sub>5</sub>) can then be easily converted  
33 to particle nitrate through gas- or aqueous-phase reactions.

34 Nitrate episodes have been mostly observed during the nighttime and have been  
35 found to be strongly related to high NO<sub>2</sub>, NH<sub>3</sub>, O<sub>3</sub> and ambient relative humidity  
36 (*Dall'Osto et al., 2009; Pathak et al., 2011; Wang et al., 2009; Wen et al., 2015*). N<sub>2</sub>O<sub>5</sub>  
37 has not been directly measured in these studies, but has been suspected to be the main  
38 cause of the nocturnal nitrate episodes through hydrolysis process. A recent study  
39 showed higher N<sub>2</sub>O<sub>5</sub> concentrations during pollution episodes in an urban  
40 environment in Beijing, demonstrating the important contribution of nighttime nitrate  
41 formation to aerosol loading (*Wang et al., 2017*). However, the dominant formation  
42 mechanisms for nitrate during daytime remain controversial. Besides the oxidation of  
43 NO<sub>2</sub> via OH (*Alexander et al., 2009; Seinfeld & Pandis, 2012*), the hydrolysis of N<sub>2</sub>O<sub>5</sub>

44 might also be important according to a multiphase chemistry box-model study in the  
45 PRD region (Xue *et al.*, 2014). However,  $\text{N}_2\text{O}_5$  is prone to dissociate back to  $\text{NO}_3$  and  
46  $\text{NO}_2$  during daytime due to its photolysis and thermal decomposition properties  
47 (Seinfeld & Pandis, 2012; Wängberg *et al.*, 1997), and thus extremely low  
48 concentrations of  $\text{NO}_3$  and  $\text{N}_2\text{O}_5$  have been commonly observed (Chang *et al.*, 2011;  
49 Stutz *et al.*, 2010).

50 Daytime nitrate episodes have been observed in Hong Kong of the PRD by the  
51 same research group, but different interpretations of formation mechanisms have been  
52 provided (Griffith *et al.*, 2015; Xue *et al.*, 2014). The episodes in summer 2011 were  
53 attributed to fast photochemistry of  $\text{HNO}_{3(\text{g})}$  under ammonium rich conditions  
54 (Griffith *et al.*, 2015) and those in winter 2009 were assumed to be caused by  
55 hydrolysis of  $\text{N}_2\text{O}_5$  (Xue *et al.*, 2014). To shed some light on this topic, an integrated  
56 field experiment was conducted in the present study measuring water-soluble  
57 inorganic ions, trace gases, particle size distribution and aerosol vertical backscatter  
58 density at high time resolution using online instruments at an urban site in Guangzhou  
59 of the PRD. In addition, vertical profiles of humidity, temperature, and wind speed  
60 and direction were obtained from the Guangzhou national weather station located 18.0  
61 km northeast of the measurement site. Potential formation mechanisms for daytime  
62 nitrate episodes were explored through in-depth data analysis.

## 63 **2. Methodology**

### 64 **2.1 Site description**

65 The measurement site is located inside the South China Institute of  
66 Environmental Science (SCIES) ( $23^\circ 07' \text{N}$ ,  $113^\circ 21' \text{E}$ ), which represents a typical  
67 urban environment in Guangzhou, south China (Tao *et al.*, 2012). The instruments  
68 were mounted on the roof of a building 50 m above ground and the data were

69 collected during 15 October to 6 November 2015.

## 70 **2.2 Online measurements of water-soluble inorganic ions and gaseous precursors**

71 Water-soluble inorganic ions ( $\text{Na}^+$ ,  $\text{NH}_4^+$ ,  $\text{K}^+$ ,  $\text{Mg}^{2+}$ ,  $\text{Ca}^{2+}$ ,  $\text{NO}_2^-$ ,  $\text{SO}_4^{2-}$ ,  $\text{NO}_3^-$ ,  $\text{Cl}^-$ ,  
72 and  $\text{F}^-$ ) and gaseous precursors ( $\text{SO}_2$ ,  $\text{HNO}_3$ ,  $\text{NH}_3$ , and  $\text{HCl}$ ) were measured by an  
73 In-situ Gas and Aerosol Composition monitoring system (IGAC, Model S-611,  
74 Machine Shop, Fortelice International Co., Ltd., Taiwan, China) with an hourly  
75 temporal resolution (Young *et al.*, 2016). The IGAC consists of a wet annular denuder  
76 (WAD), a scrub and impact aerosol collector (SCI), and an ion chromatography (IC).  
77 Gases and aerosols were passed through WAD at a flow rate of  $16.7 \text{ L min}^{-1}$ , with a  
78  $\text{PM}_{2.5}$  cyclone. The WAD is made of two concentric Pyrex glass cylinders and set up  
79 vertically to the ground. The inner walls of the annulus are wetted with dilute  $\text{H}_2\text{O}_2$   
80 solution (5 mM in ultrapure water). At the inlet of the SCI, the opening of an  
81 L-shaped water-spray nozzle is located at the centerline of the air stream and faces  
82 upstream. Ultrapure water is fed continuously to the nozzle at  $1.0 \text{ mL min}^{-1}$  and  
83 heated to about  $140 \text{ }^\circ\text{C}$ . Hot water steam is sprayed directly towards the particle-laden  
84 air to enhance mixing and particle-liquid contact. Then, aerosols are subjected to  
85 condensation growth in the remaining section of the SCI. The grown aerosols are  
86 accelerated through a conical-shaped impaction nozzle and collected on the impaction  
87 plate. At the base of the impaction plate, standard sample of LiBr solution is added  
88 continuously to the aerosol liquid sample at around  $0.1 \text{ mL min}^{-1}$ . The gas and aerosol  
89 liquid samples from the WAD and SCI are drawn separately by a pair of syringe  
90 pumps, between which one syringe collects the current sample ( $\sim 50 \text{ min}$ ) and the  
91 other injects the previous sample. The samples are then subsequently and  
92 simultaneously analyzed for anions and cations by two IC systems.

93 Cations ( $\text{Li}^+$ ,  $\text{Na}^+$ ,  $\text{NH}_4^+$ ,  $\text{K}^+$ ,  $\text{Mg}^{2+}$  and  $\text{Ca}^{2+}$ ) were measured by an IC system

94 (Dionex ICS-1100) using a CS12A column with 20 mM methanesulfonic acid eluent.  
95 Anions ( $F^-$ ,  $Br^-$ ,  $NO_2^-$ ,  $SO_4^{2-}$ ,  $NO_3^-$  and  $Cl^-$ ) were measured by another IC system  
96 (Dionex ICS-2100) using an AS19 column with 20 mM KOH eluent. System blank  
97 values were subtracted from sample concentrations. Method detection limits of ions  
98 were within the range of 0.001 to 0.003 mg L<sup>-1</sup>.

99 A known concentration of LiBr was used as internal standards by adding to the  
100 aerosol liquid samples to determine the stability of IGAC system. The changes of  
101 internal standards during the study period were within 8%. Eight concentration  
102 gradients of the mixed standard solutions containing the above-mentioned cations and  
103 anions (0.005, 0.01, 0.02, 0.05, 0.1, 0.2, 0.5 and 1.0 mg L<sup>-1</sup>) were used to calibrate the  
104 IC systems' response with respect to analytic concentrations. The calibration curves  
105 for each ion were then obtained and used to quantify the ions concentrations of  
106 ambient samples. The IC systems were calibrated with above eight sets of mixed  
107 standard solutions once every four weeks. The uncertainties of the IC systems were  
108 generally less than 15% according to the calibration curves for each ion.

### 109 **2.3 Online measurements of aerosol size**

110 A Scanning Mobility Particle Sizer (SMPS; TSI Model 3936, TSI, Inc., St. Paul,  
111 MN) combined with a long Differential Mobility Analyzer (DMA; TSI Model 3080)  
112 and a Condensation Particle Counter (CPC; TSI Model 3010) was used to measure  
113 submicron particle size. The sample and sheath flows through the long DMA were 0.3  
114 and 3.0 L min<sup>-1</sup>, respectively. It allows particles to be classified into the range of 14  
115 nm - 615 nm ( $D_{mob}$ , mobility diameter) with an inside impact inlet (nozzle size is  
116 0.0457cm). A total of 167 bins (sizes) of particles were scanned and measured in 5  
117 min. An Aerodynamics Particle Sizer (APS; TSI Model 3321) was used to measure  
118 the distributions of fine and coarse particle sizes. The sample and sheath flow rates of

119 APS were 1.0 and 4.0 L min<sup>-1</sup>, respectively. A total of 52 bins (sizes) of particles in  
120 the range of 542 nm - 10 µm ( $D_a$ , aerodynamics diameter) were measured at 5 min  
121 resolution.

122 Ambient air first passed through two total suspended particulate (TSP) cyclones,  
123 then passed through two stainless steel tube and two Nafion driers (Perma Pure,  
124 Model MD-700-36S-1) prior to be measured by the SMPS and APS. The dry  
125 sweeping gas of two Nafion driers was produced by an air compressor pump. Two  
126 relative humidity sensors were set up in the links between Nafion driers, SMPS and  
127 APS. The relative humidity of sample flow was controlled to be below 30% by the  
128 flow rate of the dry sweeping gas.

#### 129 **2.4 Online measurements of vertical backscatter density**

130 Vertical backscatter density was measured by a ceilometer (Vaisala Corp.,  
131 Helsinki, Finland; Model CL-31) (*Munkel et al., 2007*). CL-31 is equipped with the  
132 InGaAs diode laser, which emits pulses at a wavelength of 910 nm. Backscattered  
133 vertical profile of aerosol signal was observed at a temporal resolution of 2 second  
134 and spatial resolution of 15 meter. The entire range of the signal is 5.0 km height in  
135 this study.

#### 136 **2.5 Online measurements of meteorological parameters**

137 The ground meteorological parameters (e.g. wind direction and speed, relative  
138 humidity (RH), temperature, and precipitation) were measured every 10 min at an  
139 automatic meteorological station (Vaisala Company, Helsinki, Finland, model  
140 MAWS201), which was mounted on the roof of the SCIES site (53 m above ground).  
141 Vertical profiles of RH, temperature, and wind speed and direction were obtained  
142 from the Guangzhou national weather station (23°13'N, 113°29'E) located 18.0 km  
143 northeast of the SCIES site. Vertical profiles of RH and temperature were measured

144 by a high precision microwave radiometer (Radiometer Physics GmbH,  
145 RPG-HATPRO-G4, Meckenheim, Germany). The measurement range of the  
146 microwave radiometer was only up to 2000 meters height with 10-100 meters  
147 resolution. Vertical profiles of wind speed and direction were measured every 6-min  
148 by a wind profiler (Aerospace Science & Industry Corporation, CFL-16, China),  
149 which provided 25 levels of data below ~3 km with a vertical resolution of 120 m.

## 150 **2.6 Data analysis**

151 Liquid water content (LWC) shown in Fig. 1 was calculated using the  
152 ISORROPIA II thermodynamic equilibrium model for  $K^+$ - $Ca^{2+}$ - $Mg^{2+}$ - $NH_4^+$ - $Na^+$ -  
153  $SO_4^{2-}$ - $NO_3^-$ - $Cl^-$ - $H_2O$  aerosols. The input data included  $K^+$ ,  $Ca^{2+}$ ,  $Mg^{2+}$ ,  $NH_4^+$ ,  $Na^+$ ,  
154  $SO_4^{2-}$ ,  $NO_3^-$ ,  $Cl^-$ ,  $HNO_{3(g)}$ ,  $NH_{3(g)}$ ,  $HCl_{(g)}$ , ambient RH and temperature on the ground,  
155 and the forward mode was used for the calculation (*Fountoukis and Nenes, 2007*).  
156 The uncertainties of the calculated LWC were estimated to be less than 20%  
157 according to the measurement uncertainties of the water-soluble inorganic ions,  
158 gaseous precursors, ambient RH and temperature.

159 To investigate the air mass origins of the pollutants arriving at SCIES, 24 h  
160 backward trajectories were calculated at an elevation of 500 m every hour during the  
161 daytime nitrate episodes on 27 and 28 October in 2015 using the HYbrid  
162 Single-Particle Lagrangian Integrated Trajectory (HYSPLIT) 4 model  
163 (<http://ready.arl.noaa.gov/HYSPLIT.php>).

## 164 **3. Results and Discussion**

### 165 **3.1 Elevated nitrate episodes**

166 The average concentrations of  $SO_4^{2-}$ ,  $NO_3^-$  and  $NH_4^+$  were  $10.9\pm 3.5$ ,  $4.9\pm 6.5$  and  
167  $3.9\pm 2.8 \mu g m^{-3}$  during the study period, accounting for  $49\pm 10\%$ ,  $16\pm 10\%$  and  $15\pm 4\%$   
168 of the total water-soluble inorganic ions (WSII) in  $PM_{2.5}$ , respectively (Table 1).

169 Although the concentration of  $\text{NO}_3^-$  was only half of that of  $\text{SO}_4^{2-}$  on campaign  
170 average, its hourly maximum was more than doubled that of  $\text{SO}_4^{2-}$ , suggesting the  
171 much large variations of  $\text{NO}_3^-$  than  $\text{SO}_4^{2-}$  in urban Guangzhou, as was seen from their  
172 standard deviations and the 10<sup>th</sup>-90<sup>th</sup> percentile values. To explore possible  
173 mechanisms causing the high nitrate concentration, nitrate episodes were first  
174 identified as cases with hourly  $\text{NO}_3^-$  concentration above the 90<sup>th</sup> percentile ( $> 9.5 \mu\text{g}$   
175  $\text{m}^{-3}$ ) and continued for more than five hours. A total of four episodes were identified  
176 with two occurred during daytime (on 27 and 28 October) and the other two during  
177 nighttime (on 15 and 16 October) (Fig. S1).

178 The average concentrations of  $\text{NO}_3^-$  were  $17.9 \mu\text{g m}^{-3}$  and  $30.6 \mu\text{g m}^{-3}$  during the  
179 daytime and nighttime episodes, respectively (Table 1). Compared with the  
180 non-episodes,  $\text{NO}_3^-$  was enhanced the most (by a factor of 5.1 and 8.7 during daytime  
181 and nighttime episodes, respectively), followed by  $\text{NH}_4^+$  (a factor of 3.0 and 4.2) and  
182  $\text{SO}_4^{2-}$  (a factor of 1.6 and 1.8). Their gaseous precursors (e.g.  $\text{NH}_3$ ,  $\text{SO}_2$ ,  $\text{HNO}_3$ ,  $\text{NO}_x$ )  
183 were only enhanced by a factor of 1.2-1.8 during the nitrate episodes. The somewhat  
184 more enhanced  $\text{NH}_4^+$  than  $\text{SO}_4^{2-}$  was due to its association with  $\text{NO}_3^-$ . The  
185 enhancement rates of  $\text{NO}_3^-$  were much higher than those of  $\text{SO}_4^{2-}$ , especially during  
186 the nighttime episodes due to the hydrolysis of  $\text{N}_2\text{O}_5$  (Fig.S2). As expected, the  
187 average ambient RH at ground level during the nighttime episodes was close to its  
188 90<sup>th</sup> percentile value. Although the average ambient temperature at ground level  
189 during the daytime episodes was equal to its 90<sup>th</sup> percentile value, the average  
190 ambient RH was still much higher than its average and median values, suggesting the  
191 daytime episodes were likely related to aqueous-phase reactions.

192

193

*Insert Table 1*

194

195 **3.2 Formation mechanisms for daytime elevated ambient nitrate**

196 To shed some light on the possible formation mechanisms causing the daytime  
197 nitrate episodes, relevant parameters including deliquescence relative humidity  
198 (DRH), dissociation constant ( $K_p$ ) of  $\text{NH}_4\text{NO}_3$  and LWC of inorganic aerosols were  
199 calculated (Fig. 1). DRH and  $K_p$ , the reciprocal functions of ambient temperature, can  
200 be used to evaluate the thermodynamic properties of  $\text{NH}_4\text{NO}_3$ . When ambient RH >  
201 DRH, nitrate is prone to exist in aqueous phase, otherwise in solid phase (Seinfeld &  
202 Pandis, 2012). Nitrate in solid phase ( $\text{NH}_4\text{NO}_3$ ) can dissociate to  $\text{HNO}_{3(\text{g})}$  and  $\text{NH}_{3(\text{g})}$   
203 at high  $K_p$  value condition. The mole equivalent concentrations of  $\text{NH}_4^+$ ,  $\text{SO}_4^{2-}$  and  
204  $\text{NO}_3^-$  indicated that there was enough  $\text{NH}_4^+$  to neutralize both  $\text{SO}_4^{2-}$  and  $\text{NO}_3^-$  (Fig. 2).  
205 Thus,  $\text{NH}_4\text{NO}_3$  could exist in solid phase or in aqueous phase as  $\text{NH}_4^+$  and  $\text{NO}_3^-$   
206 according to the relationship between ambient RH and DRH.

207 Fig. 1 shows that the ambient RH was evidently lower than or equal to DRH and  
208  $K_p$  was at high levels during the daytime episodes.  $\text{NH}_4\text{NO}_3$  was thus expected to be  
209 existed in solid phase, and was prone to dissociate to  $\text{NH}_3$  and  $\text{HNO}_3$ . Interestingly,  
210  $\text{NO}_3^-$  concentration increased rather decreased with increasing  $K_p$  value.  $\text{OH}\cdot$  was not  
211 measured in this study, but its diurnal variation was likely similar to that of  $\text{O}_3$  and  
212 was generally at high concentration levels during the daytime in the PRD region  
213 (Hofzumahaus *et al.*, 2009). The concurrent increases of  $\text{O}_3$  and  $\text{NO}_3^-$  raised the  
214 possibility of significant contribution of  $\text{NO}_2$  oxidation via  $\text{OH}\cdot$  to daytime nitrate  
215 episode. However, simultaneous increases of nitrate and its gaseous precursors (e.g.  
216  $\text{NO}_2$  and  $\text{HNO}_3$ ) concentrations excluded such a possibility. Furthermore,  $\text{NO}_x$   
217 concentrations were slightly lower during the episodes than the average values before  
218 or after the episodes. The paradox phenomena mentioned above indicated other

219 possibilities causing the daytime episodes.

220 Particle size distribution can provide some evidences on the formation  
221 mechanisms of major inorganic ions in the atmosphere (Yao *et al.*, 2003; Zhuang *et al.*,  
222 1999). The condensation mode particles can be considered from gas-phase reactions  
223 while the droplet model from aqueous-phase processes (John *et al.*, 1990). As shown  
224 in Fig. 3a and Fig. 3b, particle number concentration in droplet mode (aerodynamic  
225 diameter ranged from 600 nm to 700 nm) evidently increased during the daytime  
226 episodes while that of condensation mode (mobility diameter < 200 nm) changed little.  
227 Aerodynamic diameters of condensation mode particles were estimated to be less than  
228 300 nm based on their average density (about  $1.5 \text{ g cm}^{-3}$ ) of the measured  
229 size-resolved chemical components on 27 and 28 October. Thus, the daytime nitrate  
230 episodes were likely from aqueous-phase processes rather than the gas-phase  
231 reactions between  $\text{NH}_3$  and  $\text{HNO}_3$ .

232

233 *Insert Fig. 1*

234 *Insert Fig. 2*

235 *Insert Fig. 3*

236

237 As expected, aerosol backscatter density near the ground showed a good  
238 agreement with the timing of daytime episodes on 27 and 28 October (Fig. 4). Clouds  
239 were identified at 500-1000 meters above the ground before daytime elevated nitrate  
240 episodes. These clouds gradually disappeared afterwards due to strong solar radiation  
241 causing cloud droplets evaporation. Aerosol particles produced from evaporated cloud  
242 droplets can be vertically transported to the surface through turbulent mixing and  
243 gravitational setting. For example, clouds appeared at about 750 meters above the

244 ground during 9:30-12:00 on 27 October (Fig. 4), dissipated afterwards, and  
245 accompanied with enhanced nitrate concentration during 12:00 to 16:00 on 27  
246 October (the first daytime nitrate episode). The time evolution of the vertical  
247 distribution of aerosol backscatter density showed vertical diffusion of the cloud  
248 evaporation produced particles both downward and upward (Fig. 4). A similar pattern  
249 was also observed during the second daytime nitrate episode (on 28 October). It is  
250 likely that cloud processes played an important role causing daytime nitrate episodes.

251

252

*Insert Fig. 4*

253

### 254 ***3.3 Dynamics mechanisms of cloud processed aerosols***

255 To further investigate the effective dynamics mechanisms of transporting cloud  
256 processed aerosols to surface-level during the daytime episodes, impact of regional  
257 transport should be first assessed. Horizontal wind speed and direction at different  
258 heights were plotted during the daytime nitrate episodes on 27 and 28 October (Fig.  
259 5a and 5b). Here, east wind is defined as wind coming from the east and the same  
260 definition is used for other wind directions. Wind directions below the 1000 m height  
261 were from the northwest to northeast during the nitrate episode on 27 October and  
262 from the southeast during the nitrate episode on 28 October. Wind speeds at these  
263 levels were relatively weak ( $< 2 \text{ m s}^{-1}$ ). Similar wind directions were also obtained  
264 from air mass back trajectory analysis during these periods. For example, the weak  
265 wind from the northwest direction at 15:00 on 27 October and the strong wind from  
266 the southeast direction at 17:00 on 28 October were simulated by back trajectory  
267 analysis (Fig. 6a and 6b), which agreed with those observed by wind profiler at the  
268 height of 200 meters (Fig. 5a and 5b). During these daytime episodes, air masses

269 mainly originated from the southeast direction of Guangzhou with paths having few  
270 significant pollutant sources except urban Hong Kong (Zheng et al., 2009). In  
271 comparison, during another period (at 11:00 on 27 October, a few hours before the  
272 first episode) with no cloud evaporation but with air mass also passing over urban  
273 Hong Kong, much lower  $\text{NO}_3^-$  concentration was observed (Fig. 6a). Furthermore, the  
274 non-episode high  $\text{NO}_3^-$  concentrations were actually mainly associated with south and  
275 southwest wind directions (Fig. 7). Thus, the daytime nitrate episodes were not caused  
276 by regional transport of pollutants.

277

278 *Insert Fig. 5*

279 *Insert Fig. 6*

280 *Insert Fig. 7*

281

282 Clouds formed about 2-3 hours before the daytime nitrate episode were at the  
283 height just above the top of boundary layer and were shallow stratocumulus clouds  
284 with a cloud depth of about 170-230 meters, which can be evaporated under strong  
285 solar radiation after sunrise. An inversion layer at the cloud top might have existed,  
286 but this cannot be firmly confirmed with the data we have. This is because the cloud  
287 top during the daytime periods was higher than 1000 meter, likely in the range of  
288 1000-2700 meters as shown in Figure 4, while the measurement range of the  
289 microwave radiometer only covered the height up to 2000 meter and with a very low  
290 resolution of 100 meter at upper levels. Thus, if the cloud top was higher than 2000  
291 meter, or if the cloud top was in between 1000-2000 meter but the inversion layer was  
292 very thin, the instrument would not be able to detect such an inversion layer.  
293 Furthermore, it is also possible that the temperature profiles were different between

294 the measure site (SCIES station) and the weather station 18 km away.

295 Despite the potential inversion layer at the cloud top during the 27 and 28  
296 October episode, no inversion was seen within the boundary layer (<2000 meters)  
297 (Fig. 8a), providing the needed meteorological conditions for transporting residual  
298 aerosols produced from cloud evaporation to the lower-levels through strong turbulent  
299 mixing. Strong turbulence was indeed observed during the daytime nitrate episodes  
300 with the vertical velocity up to  $0.3 \text{ m s}^{-1}$  at many levels below the original cloud base  
301 (Fig. 8b), enabling the residual aerosols to be able to transport the surface layer in just  
302 half to one hour. Furthermore, the transient high RH was simultaneously observed at  
303 all levels during the daytime episodes, further supporting the above hypothesis (Fig.  
304 8c). It is also noticed that the duration was longer and the average nitrate  
305 concentration was higher on 28 October than 27 October episode, likely due to the  
306 stronger vertical turbulence (as seen from the observed vertical wind speed) on 28  
307 October.

308

309 *Insert Fig. 8*

310

311 The average ratio of  $\text{NO}_3^-/\text{SO}_4^{2-}$  in ambient aerosols was  $0.40 \pm 0.37$  during the  
312 study period, which was within the range of the ratio (0.3-0.6) in precipitation in  
313 China during 2001-2015 (Itahashi *et al.*, 2017). However, the daytime nitrate episodes  
314 showed a much higher  $\text{NO}_3^-/\text{SO}_4^{2-}$  ratio (1.05), which was likely due to recycling of  
315 cloud condensation nuclei with each cycle having added nitrate. Furthermore, the  
316 moist condition in cloud was in favor of the reserve of  $\text{NO}_3^-$  in aqueous phase and  
317 prevented it from dissociating into  $\text{HNO}_{3(\text{g})}$  and  $\text{NH}_{3(\text{g})}$ . In conclusion, vertical  
318 turbulent mixing of the residual aerosols produced from cloud evaporation well

319 explained the daytime episodes and the high concentrations of the droplet mode  
320 particles.

#### 321 **4. Summary and Implication**

322 Differing from existing studies which frequently attributed high surface-level  
323 aerosol concentrations to enhanced source emissions and secondary aerosol formation,  
324 the present study identified a potentially important mechanism causing high nitrate  
325 concentration during daytime in the PRD. A conceptual framework is firstly needed  
326 for designing and simulating the cloud process in order to quantify its impact on  
327 nitrate concentration budget on seasonal and annual time scales. This framework  
328 should include aerosol-cloud interaction, gas- and aqueous-phase chemistry involving  
329  $\text{NH}_4^+$ ,  $\text{SO}_4^{2-}$  and  $\text{NO}_3^-$  particles, and more importantly, cloud evaporation releasing  
330 processed aerosols which can provide a pathway reasonably explaining daytime  
331 nitrate episodes in cloudy regions such as the PRD. Existing size- and  
332 chemically-resolved cloud microphysics models such as the one described in *Zhang et*  
333 *al.* (2006) can be modified for conducting such modeling studies.

334 This study suggested that  $\text{SO}_4^{2-}$  and  $\text{NO}_3^-$  were mostly fully neutralized by  $\text{NH}_4^+$   
335 in the PRD. In this scenario  $\text{NO}_3^-$  formed in cloud water would combine with  $\text{NH}_4^+$  or  
336 react with  $\text{NH}_3$  to form  $\text{NH}_4\text{NO}_3$  in solid phase, rather than dissociate to  $\text{HNO}_3$  during  
337 cloud evaporation process. With further reduced  $\text{SO}_2$  emission in many parts of China,  
338 the contribution of nitrate to fine particles become more important in ammonia rich  
339 conditions. In the PRD, emissions of  $\text{NO}_x$  from industry have declined while those  
340 from traffic have slightly increased due to the increase of the number of vehicles  
341 (*Zheng et al.*, 2009). Little attention has been paid on  $\text{NH}_3$  emission in the PRD as  
342 well as other parts of China. Emission reduction of both  $\text{NO}_x$  and  $\text{NH}_3$  should be  
343 considered together for reducing fine particles.

344 **Acknowledgments**

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**Table 1 Statistics of hourly mass concentrations of water-soluble inorganic ions in PM<sub>2.5</sub>, trace gases and selected meteorological factors in urban Guangzhou, PRD.**

	Average $\pm$ standard deviation	Minimum	10th	median	90th	Maximum	Daytime episode	Daytime non-episode	Nighttime episode	Nighttime non-episode
NH <sub>4</sub> <sup>+</sup> /μg m <sup>-3</sup>	3.9±2.8	<0.1	1.6	3.3	6.9	18.8	10.2	3.4	13.3	3.2
Na <sup>+</sup> /μg m <sup>-3</sup>	0.6±0.2	0.4	0.5	0.6	0.8	1.7	0.8	0.6	0.7	0.6
K <sup>+</sup> /μg m <sup>-3</sup>	2.0±0.3	1.5	1.6	1.9	2.4	3.6	2.0	2.0	3.2	2.0
Mg <sup>2+</sup> /μg m <sup>-3</sup>	0.2±0.0	0.2	0.2	0.2	0.2	0.3	0.2	0.2	0.2	0.2
Ca <sup>2+</sup> /μg m <sup>-3</sup>	0.1±0.1	0.0	0.0	0.1	0.2	0.7	0.1	0.1	0.1	0.1
SO <sub>4</sub> <sup>2-</sup> /μg m <sup>-3</sup>	10.9±3.5	1.8	7.0	10.7	15.3	21.2	17.1	10.8	18.6	10.1
NO <sub>3</sub> <sup>-</sup> /μg m <sup>-3</sup>	4.9±6.5	0.3	1.0	2.5	9.5	51.5	17.9	3.5	30.6	3.5
Cl <sup>-</sup> /μg m <sup>-3</sup>	0.6±0.6	0.0	0.2	0.4	1.4	5.6	1.2	0.5	1.7	0.5
NO <sub>2</sub> <sup>-</sup> /μg m <sup>-3</sup>	0.8±0.4	0.0	0.3	0.7	1.2	2.8	1.0	0.6	1.6	0.8
WSII/μg m <sup>-3</sup>	24.0±12.7	4.8	14.1	21.6	35.1	100.3	50.6	22.0	70.0	21.0
NH <sub>3</sub> /μg m <sup>-3</sup>	9.3±3.8	3.3	4.7	8.7	14.8	19.9	15.4	9.4	10.4	8.6
SO <sub>2</sub> /μg m <sup>-3</sup>	11.5±4.5	3.5	6.2	11.1	16.8	26.8	17.9	11.8	14.9	10.7
HNO <sub>3</sub> /μg m <sup>-3</sup>	1.7±0.7	0.5	1.0	1.5	2.7	5.0	3.5	1.9	2.0	1.4
NO/μg m <sup>-3</sup>	16±27	<1	1	5	44	263	10	12	41	18
NO <sub>2</sub> /μg m <sup>-3</sup>	71±35	12	33	66	122	179	84	57	109	82
NO <sub>x</sub> /μg m <sup>-3</sup>	87±54	12	35	72	155	442	94	69	150	100
HCl/μg m <sup>-3</sup>	0.1±0.2	0.0	0.0	0.0	0.3	0.8	0.4	0.2	0.0	0.0
HNO <sub>2</sub> /μg m <sup>-3</sup>	3.6±2.6	0.5	1.2	2.8	7.3	13.7	3.7	2.5	8.2	1.4
O <sub>3</sub> /μg m <sup>-3</sup>	53±52	1	3	31	134	252	114	81	46	23
RH/%	53±14	26	34	54	70	76	59	48	67	57
Temp/°C	25±3	16	21	26	29	31	29	26	25	24
WS/m s <sup>-1</sup>	1.8±1.2	0.0	0.3	1.8	3.4	5.3	1.6	2.0	0.7	1.8

WSII: water-soluble inorganic ions; RH: relative humidity; Temp: temperature; WS: wind speed; daytime: 7:00-18:00; nighttime: 19:00-6:00

**List of Figures**

Fig. 1 Diurnal variations of hourly nitrate concentrations and relevant parameters during the daytime episodes.

Fig. 2 Scatter plots of  $\text{SO}_4^{2-}$  versus  $\text{NH}_4^+$  (a),  $\text{NO}_3^-$  versus  $\text{NH}_4^+$  (b), and the sum of  $\text{SO}_4^{2-}$  and  $\text{NO}_3^-$  versus  $\text{NH}_4^+$  (c).

Fig. 3 Aerosol size distributions during the daytime episodes.

Fig. 4 Vertical distribution of aerosol backscatter densities during the daytime episodes.

Fig. 5 Vertical profiles of horizontal wind speed and direction during the daytime episodes.

Fig. 6 72 h air mass back trajectories arriving at 200 m elevation at the SCIES before (black line) and during (red line) the daytime episodes on 27 (a) and 28 October (b).

Fig. 7 Wind map of nitrate concentration during the daytime non-episode periods.

Fig. 8 Vertical profiles of temperature (a), vertical wind speed (b), and RH (c) during the daytime episodes.

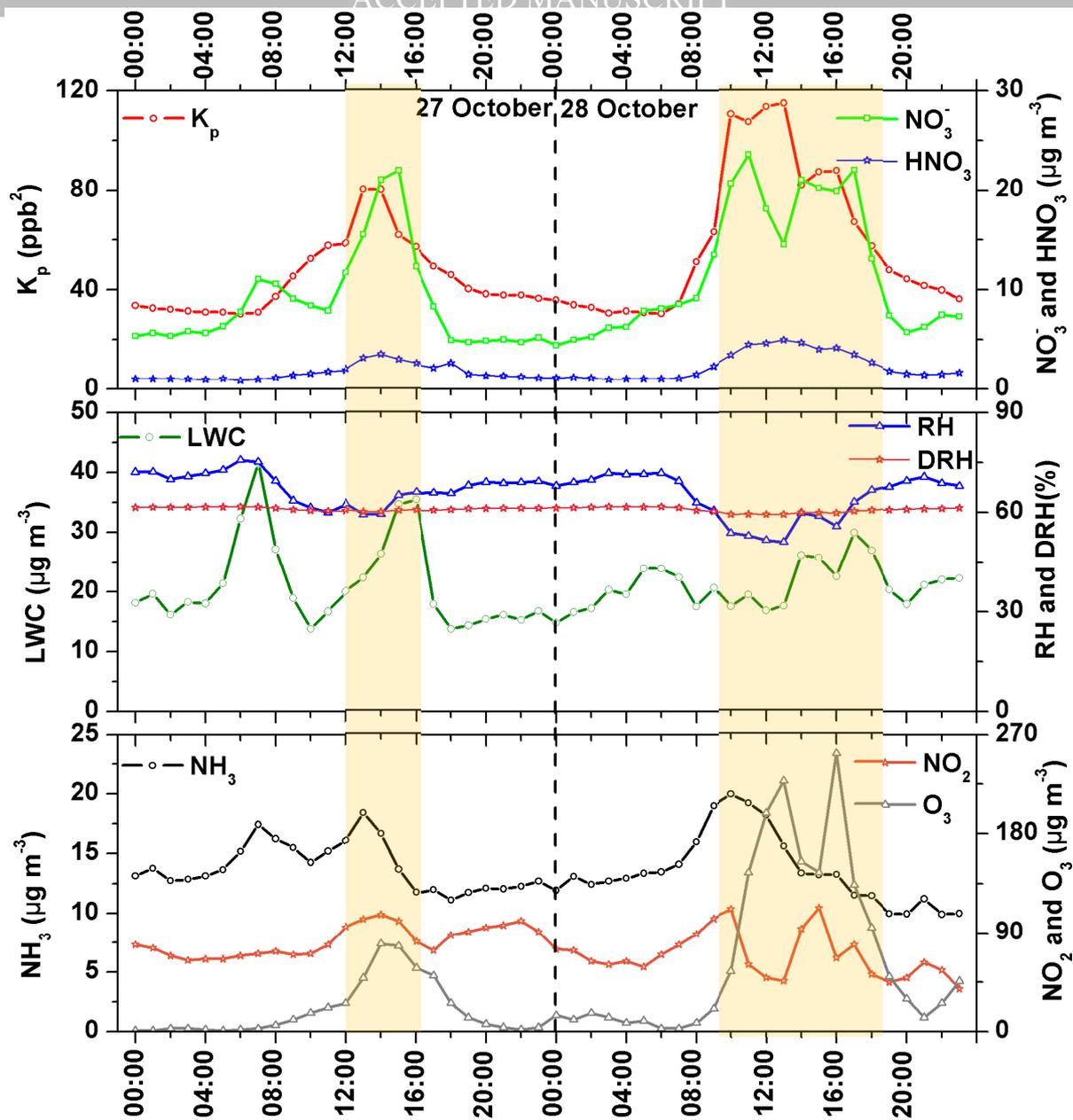


Fig. 1.

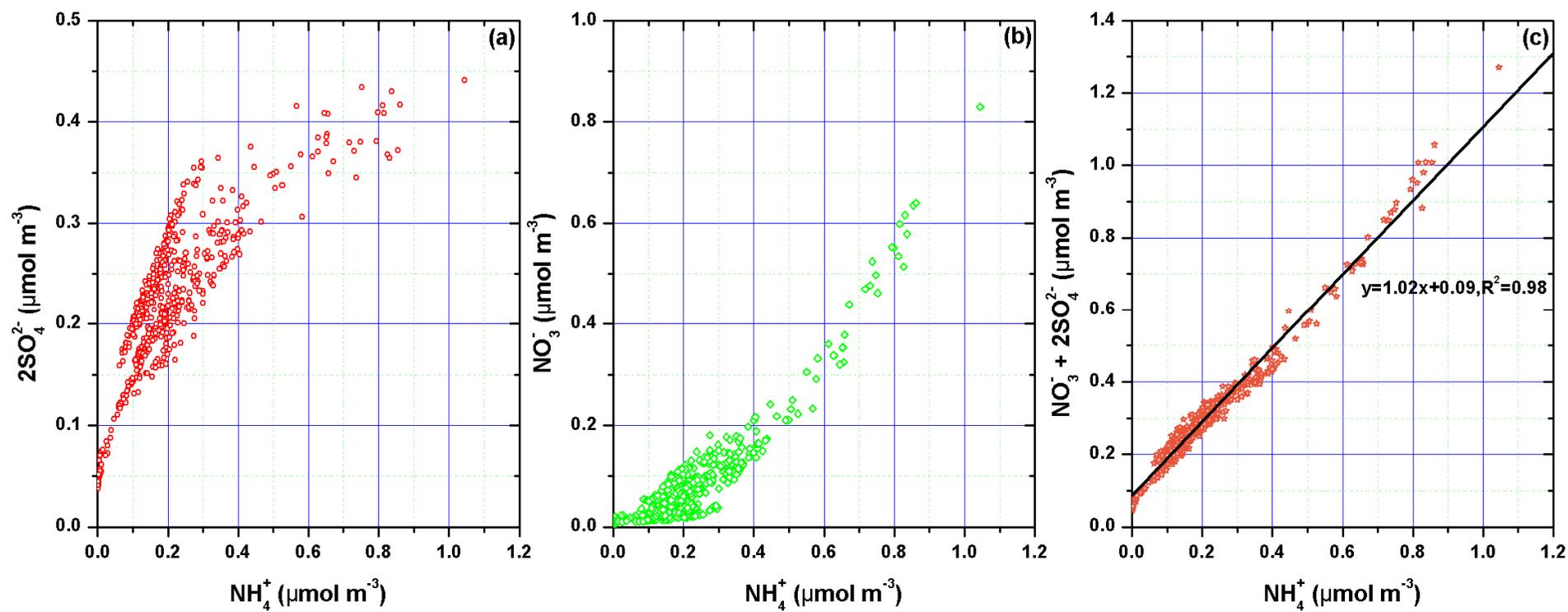
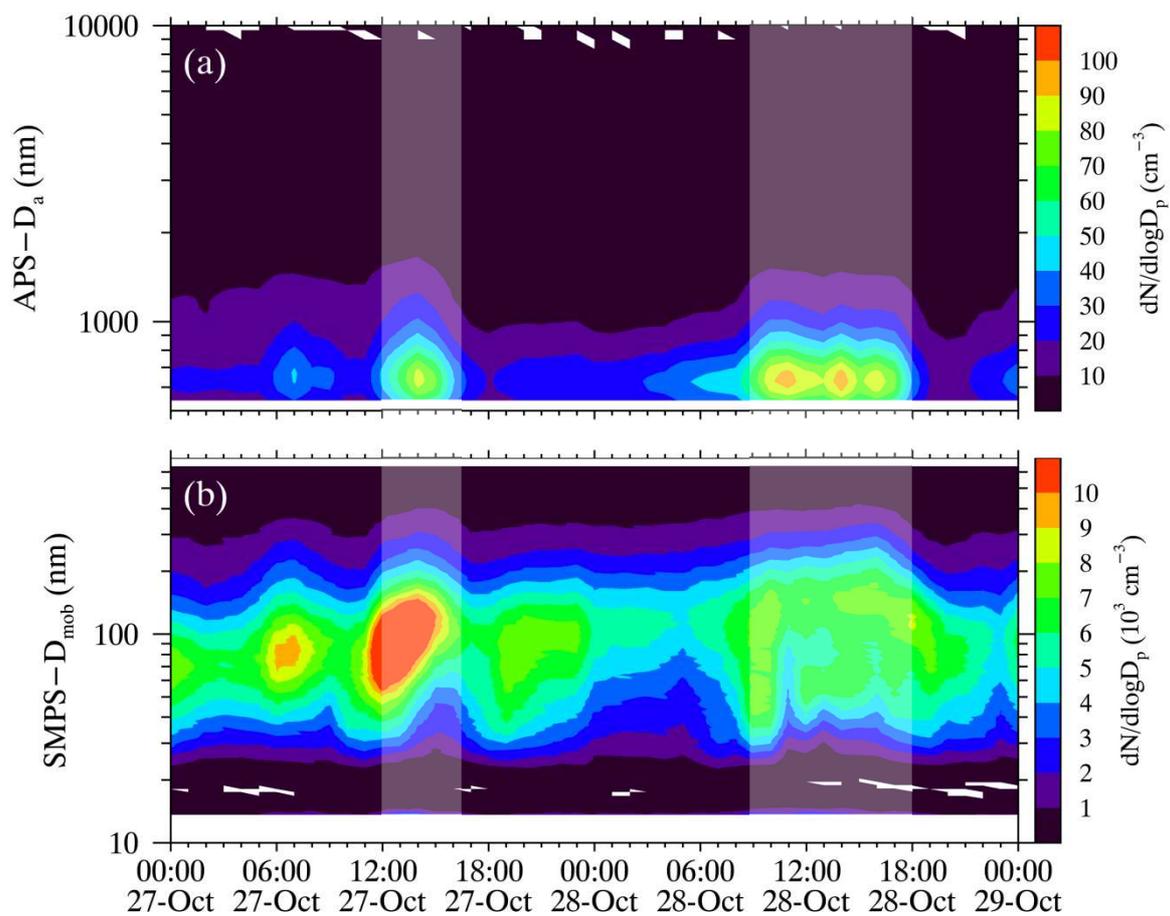


Fig. 2.

**Fig. 3.**

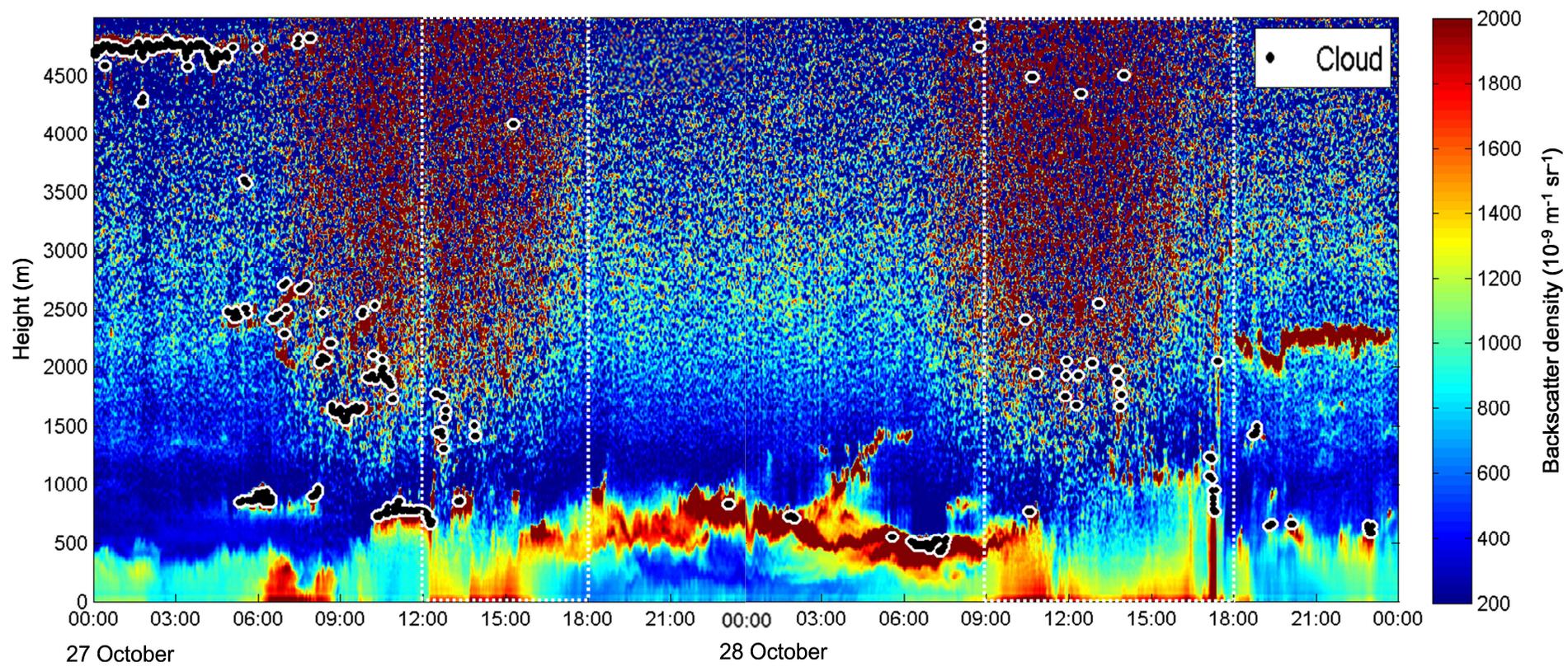


Fig. 4.

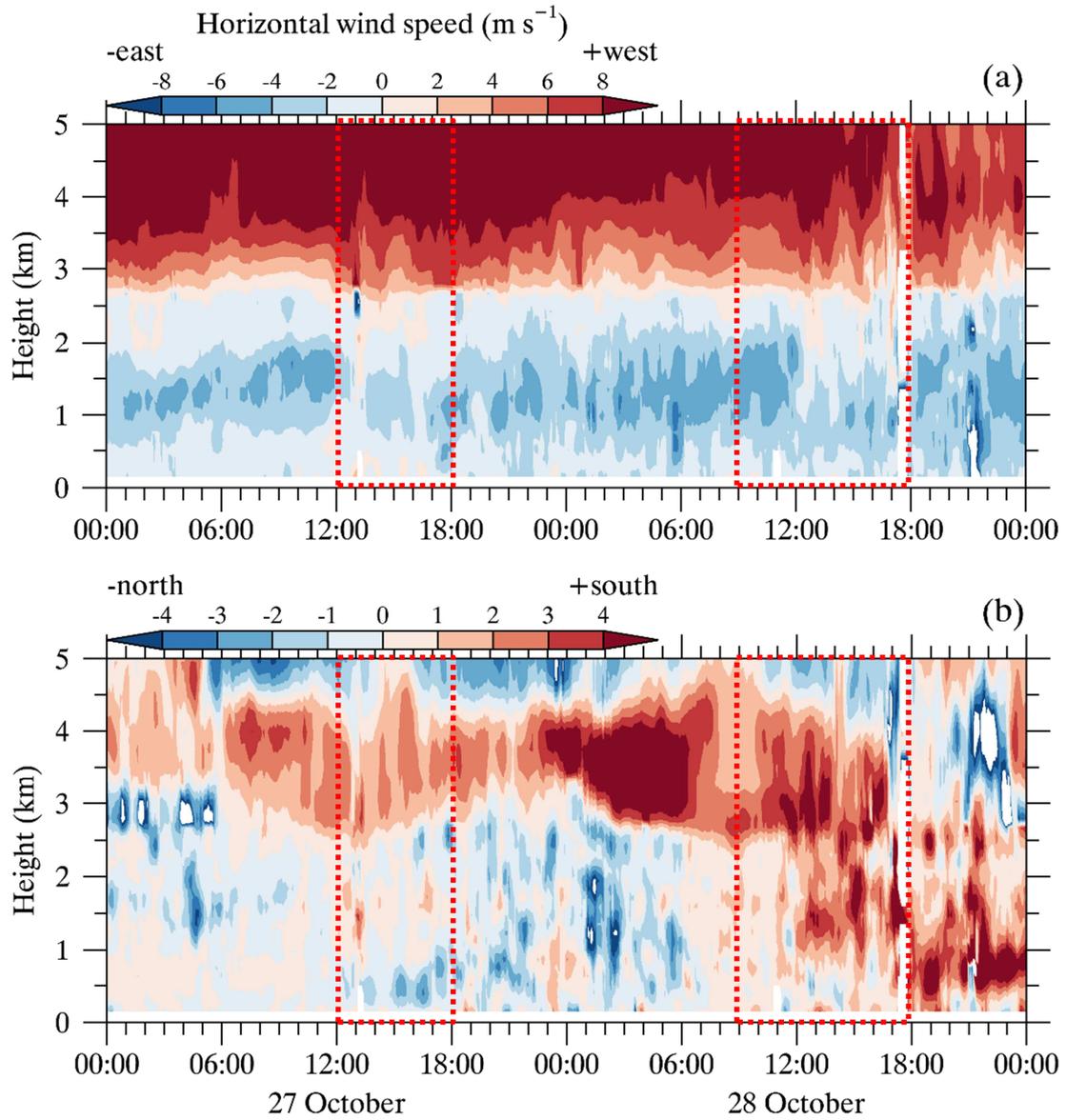


Fig. 5.

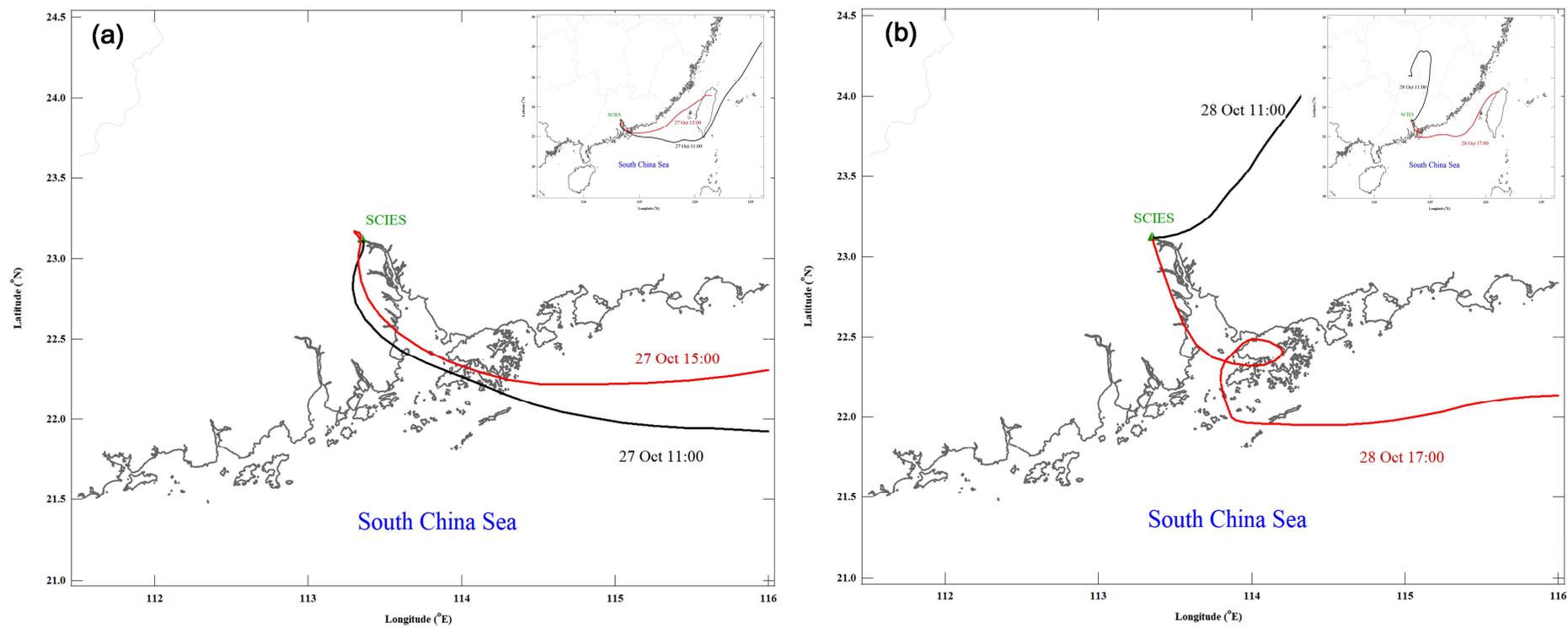


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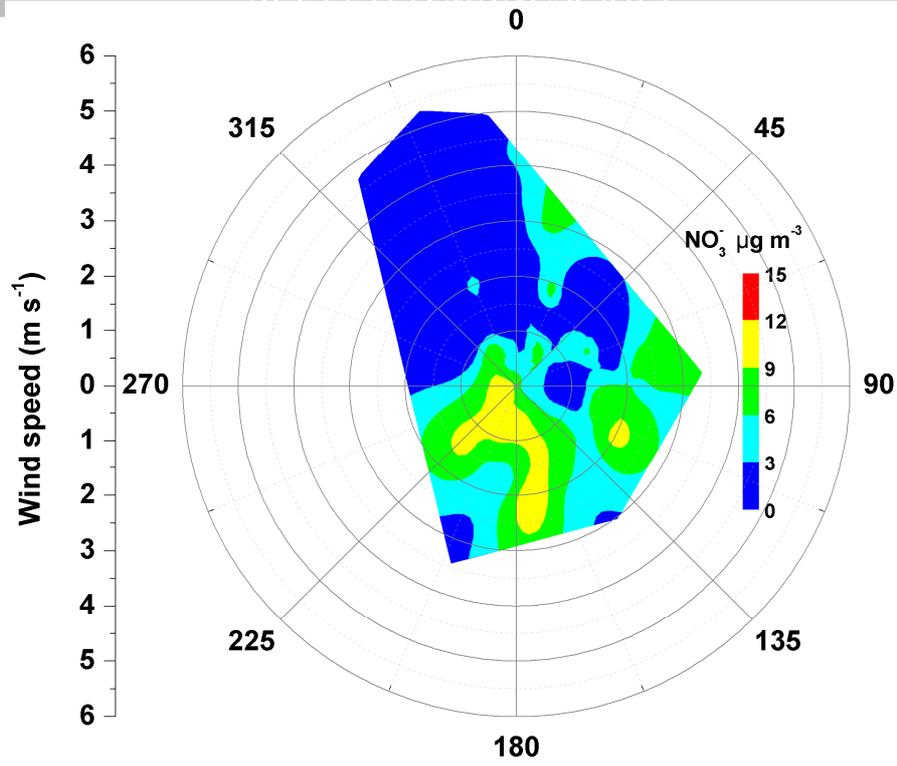


Fig. 7.

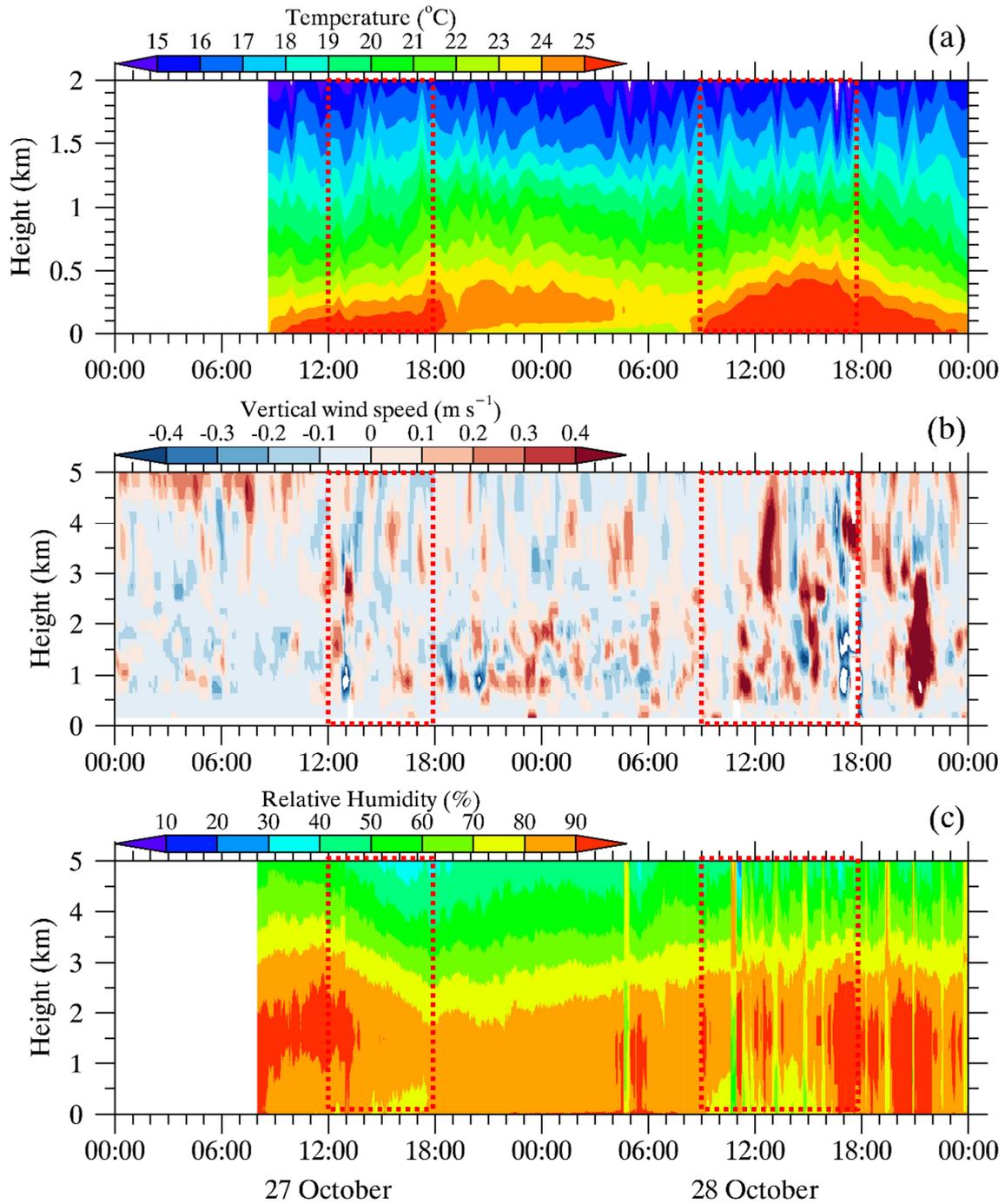


Fig. 8.

**Highlights:**

- $PM_{2.5}$  chemical composition and size distribution were measured
- Day- and nighttime nitrate episodes were observed
- Formation mechanisms causing daytime nitrate episodes were explored in detail
- Turbulent mixing of cloud processed aerosols caused daytime nitrate episodes